

# A method for Estimating the Degree of Uncertainty With Respect to Life Cycle Assessment During Design

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*Life cycle assessment (LCA) is used to estimate a product's environmental impact. Using LCA during the earlier stages of design may produce erroneous results since information available on the product's lifecycle is typically incomplete at these stages. The resulting uncertainty must be accounted for in the decision-making process. This paper proposes a method for estimating the environmental impact of a product's life cycle and the associated degree of uncertainty of that impact using information generated during the design process. Total impact is estimated based on aggregation of individual product life cycle processes impacts. Uncertainty estimation is based on assessing the mismatch between the information required and the information available about the product life cycle in each uncertainty category, as well as their integration. The method is evaluated using pre-defined scenarios with varying uncertainty. [DOI: 10.1115/1.4002163]*

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## 1 Introduction

The ratio of product mass to waste mass produced as a result of the product during its life cycle is about one to twenty [1]. These wastes are produced in each phase of the product life cycle from raw material extraction to product retirement. Sustainable development is defined in Ref. [2] as "development which meets today's needs without placing the ability of future generations to meet their needs at risk." For such development, design can play a major role [3] where major requirements for the design including those for sustainable development must be identified and satisfied throughout the process [4] as decisions taken in design affect all stages of product development [5] and in turn all phases of the product's life cycle.

Life cycle assessment [6] is currently the most promising and scientifically defensible methodology for estimating environmental impacts of a product lifecycle [7]. Currently, detailed LCA [6] is critically dependent on high volumes of product-specific data, time consuming, often unaffordable, and reliably used only after detailed design. Abridged LCA [8,9] is either incomplete or inaccurate or requires prior knowledge of what data are important [10]. There is substantial uncertainty involved in the environmental impact calculations in LCA [11]. Literature [11] stresses that estimation of impact must be accompanied by estimation of its uncertainty or imprecision without which the decisions based on these results could be misleading.

If LCA is to be used throughout the design process, the degree of uncertainty involved in the estimations must be assessed and taken into account in the decision making processes during design without which the decisions might be unduly biased or incorrect. There is a need to understand the information required for using LCA in design and the information available at each design stage to ascertain the extent to which LCA could be used at each stage of design.

The objectives of this paper are as follows.

- Understand uncertainty in the context of product lifecycle information in various stages of design. This is done using literature review and descriptive studies.
- Develop a method for estimating lifecycle environmental impacts of a product and the degree of uncertainty associated with this estimation. This is done by developing a method that integrates interval algebra [12] and weighted objectives [13] and evaluating this by using example scenarios of varying uncertainty.

## 2 Literature Review and Descriptive Studies

**2.1 Literature Review.** From a survey [14] of LCA studies, it is identified that LCA results are subject to various sources of uncertainty: uncertainties introduced by the data and the methodology such as the lack of site-specific data and the aggregation of data over different spatial and temporal scales. Studies done on finding problems with LCA argue that LCA should include an explanation of the uncertainties that arise during LCA. Uncertainty assessment is necessary for better decision support, transparency and quality comparison. However, usually this is not carried out in LCA studies due to the additional effort needed and the lack of methods [7].

The methods, e.g., Refs. [17,18], have been developed for estimating impacts, taking into account uncertainties in lifecycle inventory data (LCI) in a specific domain. Their authors argue that fuzzy intervals and numbers are more informative and closer to human judgments and perceptions than crisp numbers, thus, improving the pertinence and the interpretation of the results. Some databases have statistical distributions of data [19], which can be used in LCA for impact calculations [20]. It is emphasized that interpretation of uncertainty in data and results is an indispensable part of sound decision making and should be an integral part of the analysis itself. Tools like Simapro7 [21] and KCL-ECO [22] have some limited lifecycle inventory with data distributions, and a limited facility for uncertainty analysis based on the Monte Carlo method [23], which uses inventory values for which the distribution is available (like range, triangular, normal, or lognormal); the calculation is performed for a specified number of times, each time taking a random value within the distribution. The variation in results can be displayed in different distributions or as

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74 average or best estimate. However, this analysis is limited to es-  
 75 timating uncertainty in LCI data if the distributions for the data  
 76 are available. It cannot deal with uncertainty arising from the  
 77 design process such as those associated with the product structure  
 78 or the lifecycle phases.  
 79 Normally, probability distributions [24] are used to represent  
 80 random variability in input parameters, upper and lower bounds or  
 81 fuzzy intervals are used to represent vagueness, and sensitivity  
 82 analysis is used for methodological choices [25]. Some [26] sug-  
 83 gest that for better decision making, all types of uncertainty must  
 84 be propagated into a single result, using combined models for  
 85 simulation and approximation.  
 86 Geographical, temporal or technological differences are typical  
 87 sources for uncertainty associated with inventory data in LCA; for  
 88 instance, geographical and technological differences in life cycle  
 89 inventory data are shown to be major sources of uncertainty in  
 90 LCA for processes in waste incinerators [27,28]. In Ref. [29],  
 91 specific rules of thumb are suggested for the individual impact  
 92 categories of global warming, acidification, eutrophication, and  
 93 photo-oxidant creation; the rules quantify the difference in impact  
 94 scores necessary for it to be significant in product comparison.  
 95 The authors suggest that LCI data providers should supply quan-  
 96 titative uncertainty information, including correlation estimates  
 97 for individual parameters. Some [30] emphasize the need for a  
 98 framework for modeling data uncertainty in LCI. They [30] take  
 99 uncertainty as data inaccuracy and lack of specific data, divide the  
 100 latter into complete lack of data and lack of representative data,  
 AQ: 101 and suggest as important the parameters that cause a larger spread  
 #1 102 in the model outcome.  
 103 In Ref. [26], a method is proposed for propagation of data un-  
 104 certainty into the overall results of the LCA; it combines approxi-  
 105 mation formulae such as Gauss, Bader-Baccini, and Monte Carlo  
 106 simulation to estimate the uncertainty. In Ref. [31], it is illustrated  
 107 that in the initial stages of design, functional parameters, which  
 108 are functional requirements and constraints for the design prob-  
 109 lem, should be made available for estimating environmental im-  
 110 pacts of the design; use of statistical and sensitivity analysis are  
 111 suggested for representing uncertainty.  
 112 According to literature [11,32], uncertainty exists in LCA be-  
 113 cause of data inaccuracy, data gaps, model uncertainties, choices,  
 114 spatial and temporal variability, variability between sources, etc.  
 115 In Ref. [11], it is argued that LCA results are usually presented as  
 116 point estimates, which strongly overestimate the reliability; it is  
 117 suggested that uncertainty arise due to lack of knowledge about  
 118 the true value of a quantity. Also, stressed is the need for estimat-  
 119 ing and expressing the uncertainty. Even though there are various  
 120 available methods for performing uncertainty estimations, such as  
 121 classical statistical analysis, Bayesian statistical analysis (which  
 122 needs expert judgments to ascertain the nature of distributions),  
 123 interval algebra, vague error interval calculations, and probabilis-  
 124 tic simulation (which involves the difficult task of finding all pos-  
 125 sible events), there is still need for a framework that explicate the  
 126 important aspects of data quality and uncertainty in LCA to the  
 127 practitioner [11].  
 128 **2.2 Descriptive Studies.** We have conducted a series of de-  
 129 sign exercises and analyzed their proceedings in order to under-  
 130 stand the evolving levels of uncertainty in product lifecycle infor-  
 131 mation during design. The goal was to identify the types of  
 132 uncertainty that emerge when LCA is used in design; since this  
 133 information was not available in existing literature, we carried out  
 134 our own descriptive studies to identify these. The following is a  
 135 summary of the descriptive studies, for details see Ref. [33].  
 136 Twenty-four design exercises were conducted involving 8 design-  
 137 ers and 3 design problems; each problem was solved by each  
 138 designer using one of the three interventions—use of general de-  
 139 sign literature, use of environmentally friendly design (EFD) lit-  
 140 erature, or use of detailed impact assessment software. The de-  
 141 signers followed the "think-aloud" protocol while designing; the  
 142 whole process was videotaped and transcribed, which along with

documentations were used for protocol analysis. Out of the 24 143  
 exercises, the 16 exercises that used EFD literature and detailed 144  
 impact assessment software as intervention have been analyzed, 145  
 and the summary of results are presented below. 146

- During design of product lifecycles in each of these exer- 147  
 cises, it was observed that the structure of the product (as- 148  
 sembles, subassemblies, parts, interfaces, and features) 149  
 evolved as design progressed. 150
- Various designers considered different lifecycle phases at 151  
 different stages of their design process, each at different 152  
 levels of completeness. 153
- Designers did not necessarily consider all lifecycle pro- 154  
 cesses for each life cycle phase; in some cases these became 155  
 more comprehensive as design progressed. 156
- In some of the design exercises, designers looked for spe- 157  
 cific data on environmental impacts, which were not avail- 158  
 able in the databases accessed. 159

**2.3 Summary.** As seen in Sec. 2.1, most of the literature in 160  
 this area has been focused on identifying uncertainty associated 161  
 with LCI data [13–19,22,23] with some focus on methodology 162  
 [20,24]. However, the analysis of descriptive studies (Sec. 2.2) 163  
 illustrate that information about the lifecycle of a product contin- 164  
 ues to evolve during its development: there is evolving uncer- 165  
 tainty also in the product structure, in the completeness of the 166  
 lifecycle phases, and in the lifecycle processes considered. 167  
 Traditionally, LCA is used after the detail design when detailed 168  
 information about the product, its lifecycle phases, and associated 169  
 data are available. In this case, the uncertainty will be confined to 170  
 data and methodology, depending on the variations in these. How- 171  
 ever, if LCA is used during earlier stages of design where infor- 172  
 mation about the product and its lifecycle phases are also uncer- 173  
 tain, there is a greater degree of uncertainty. Hence, in these 174  
 phases it is important to consider reducible uncertainties like those 175  
 associated with product structure and lifecycle phase along with 176  
 data and methodological uncertainty. For decision-making, the re- 177  
 sults should encompass both impact and associated uncertainty. 178  
 While literature discusses uncertainty of impact data, there is no 179  
 discussion on how to calculate and represent the overall uncer- 180  
 tainty in the estimated potential impact of a product lifecycle pro- 181  
 posal at any given stage in design with respect to LCA. 182  
 Therefore, a method for assessing environmental impacts for 183  
 product life cycles should not only provide an estimate of the 184  
 impact but also the associated degree of uncertainty that takes into 185  
 account all these various sources of uncertainty. 186  
 The following section details the uncertainty categories identi- 187  
 fied in our work from literature and descriptive studies. 188

**3 Uncertainty Categories** 189

While existing literature discusses uncertainty in data and meth- 190  
 odology; analyses of descriptive studies identified further uncer- 191  
 tainty in product structure and life cycle phases. Impact estimation 192  
 requires two things: the data and the methodology to process the 193  
 data. The data pertain to processes related to the various elements 194  
 of the product in its various lifecycle phases. Therefore, the over- 195  
 all uncertainty is affected by the uncertainty related to the product, 196  
 its life cycle phases, and those related to the data pertaining to the 197  
 processes and the methodology used to integrate this data. There- 198  
 fore, in the context of LCA, these four are the only possible ele- 199  
 ments of uncertainty. We take uncertainty as the accuracy of the 200  
 estimation rather than the probability of finding the correct esti- 201  
 mate. The four uncertainty categories are further elaborated be- 202  
 low. 203

**3.1 Product Structure.** Uncertainty about the structure of a 204  
 product is related to its subsystems, parts and interfaces. LCA 205  
 requires information about the materials and processes used in the 206  
 life cycle of the product. A product's structure fundamentally con- 207  
 tains only parts and interfaces, each having various features. 208

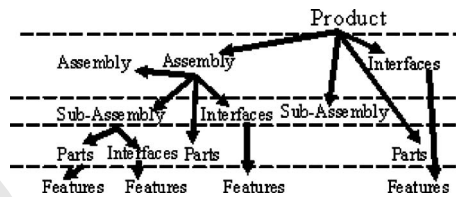


Fig. 1 A product and its subsystems

209 These parts and interfaces are hierarchically organized into groups  
 210 called assemblies and subassemblies, where subassemblies con-  
 211 tain only parts and interfaces while assemblies also contain sub-  
 212 assemblies or other assemblies. The organization is important for  
 213 capturing information about the various lifecycle processes, e.g.,  
 214 an assembly process that requires movement of the subassembly  
 215 as a whole and not as its individual parts and features. The cat-  
 216 egories (Fig. 1) provide a complete set for describing a product's  
 217 structure and are important for identification of the life cycle pro-  
 218 cesses associated with the product. For instance, while material  
 219 choice depends only on individual parts, manufacturing processes  
 220 are dependent on part features, and assembly processes depend on  
 221 the interfaces between features belonging to different parts, which  
 222 may belong to different subassemblies or assemblies. Also, these  
 223 categories are standard categories used in describing CAD mod-  
 224 els, such as in CATIA [34], and are important to be so, since a  
 225 designer would typically use a CAD model for developing and  
 226 describing a product's structure, which is required for defining the  
 227 product's life cycle processes. Uncertainty in product structure  
 228 definition is subdivided into the following (qualitative degrees of  
 229 each uncertainty are proposed within brackets).

- 230 • Uncertainty in definition of assemblies, i.e., the collection of  
 231 assemblies, subassemblies, parts, and interfaces between  
 232 them in that particular assembly of the product (all, some,  
 233 none).
- 234 • Uncertainty in definition of subassemblies, i.e., the collec-  
 235 tion of parts and interfaces in the subassemblies of the prod-  
 236 uct (all, some, none).
- 237 • Uncertainty in definition of interfaces, i.e., the connection  
 238 between one or more features of one part and one or more  
 239 features of another part in the product (all, some, none).
- 240 • Uncertainty in definition of parts, i.e., the smallest physical  
 241 element in the product, not in size but in that it cannot be  
 242 divided further into parts and interfaces (all, some, none).
- 243 • Uncertainty in definition of features, i.e., the geometrical  
 244 forms in a part (all, some, none).

245 **3.2 Lifecycle Phases.** This uncertainty is related to the mate-  
 246 rial, production, distribution, usage, and after-use phases of the

product life cycle. There are also subphases within each of these: 247  
 extraction, manufacturing, and transportation in the material phase 248  
 (all, some, none); manufacturing and assembly in the production 249  
 phase (all, some, none); packaging and transportation in the dis- 250  
 tribution phase (all, some, none); use, maintenance, and repair in 251  
 the usage phase (all, some, none); and reuse, recycle, and disposal 252  
 in the after-use phase (all, some, none). For example, at a particu- 253  
 lar stage of design, a designer may have information only about 254  
 the material of a component, and not about its other phases. The 255  
 uncertainty in the lifecycle-phases category is accounts for 256  
 whether or not a designer considers individual phases (i.e., mate- 257  
 rial, production, distribution, usage, or after-usage). Table 1 shows 258  
 some instances of designer utterances, from an exercise from the 259  
 descriptive studies, that involves the lifecycle phases of the prod- 260  
 uct. Note that many of these deliberations involve classes (non- 261  
 crisp) of lifecycle processes—such as plastics rather than a specific 262  
 plastic, transportation rather than transportation by a specific 263  
 means, etc. These would affect the specificity of values and asso- 264  
 ciated uncertainty. 265

**3.3 Data Quality.** This uncertainty is related to the relevance 266  
 of data in terms of its temporal relevance, spatial relevance, and 267  
 sample size, see details below. The uncertainty in the data quality 268  
 category can be in terms of the data being old (temporal), nonlocal 269  
 (spatial) and the number of sources on which the data are based 270  
 (sample size). Uncertainty in data quality is subdivided into the 271  
 following. 272

- Uncertainty in temporal relevance of the data (current, old, 273  
 very old): how close in time the data collected is to when the 274  
 process it describes is to be used. 275
- Uncertainty in spatial relevance of the data (national, conti- 276  
 nental, world): geographically how close the area from 277  
 which the data collected is to where the process it describes 278  
 is to be used. 279
- Uncertainty in sample size on which the data is based (mul- 280  
 tiple samples, single sample): in terms of the number of 281  
 samples used for creating the data. 282

**3.4 Methodological Choices.** This uncertainty comes from 283  
 the temporal relevance, spatial relevance and the comprehensiveness 284  
 of the methodology. The uncertainty in methodological 285  
 choices can be in terms of being old, being from a different region 286  
 than where applied, and in terms of only some of the potential 287  
 impacts being considered. Uncertainty in methodological choices 288  
 is subdivided into the following. 289

- Uncertainty in temporal relevance of the choices: how re- 290  
 cent (current, old, very old). 291
- Uncertainty in spatial relevance of the choices: how close 292  
 geographically (national, continental, world). 293

Table 1 Lifecycle processes and protocol instances from a design exercise

Lifecycle process	Protocol instance
Material	Balloons in terms of rubber, plastic, flexible material probably cloth I can use those ( <i>designer trying to evaluate and select material</i> )
Production	Will be injection molded; Mainly stitching and aluminum frame bolted ( <i>designer trying to select the production (manufacturing and assembly) processes required for the solution</i> )
Distribution	It should be easy to pack, no damage in transportation ( <i>designer is generating the requirements for product's distribution phase</i> )
Usage	It should not have any maintenance ( <i>designer generating requirement for usage phase</i> )
After-usage	Easy to disassemble; should be recyclable ( <i>designer generating the requirement of after-usage phase for the solution</i> )



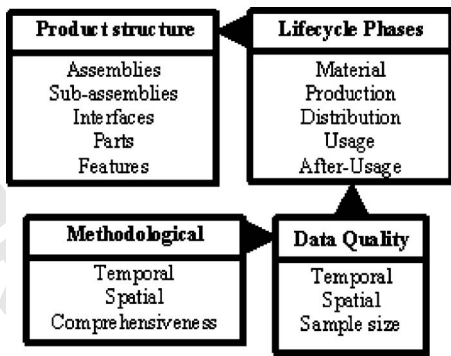


Fig. 2 Uncertainty propagation

- Uncertainty in comprehensiveness of the choices: how comprehensive the categories of impact considered by the methodology are (all, some, none).

**3.5 Uncertainty Propagation.** Figure 2 shows the uncertainties in different categories and their propagation to the overall uncertainty.

There can be uncertainty in the product structure—i.e., the definition of the product is uncertain. For instance, take a product that has two parts, Part1, Part2, and one interface Int1 between these parts such as a cutting edge connected to a handle for a vegetable cutting knife. If information about the interface is not available, e.g., how the handle is connected to the cutting edge is yet to be defined, there would be uncertainty in the product structure definition. Even if the product structure definition is complete, there can still be uncertainty in terms of the definition of the product lifecycle. For instance, the after-usage details of Part1 and Part2 may not be specified yet, giving uncertainty in the lifecycle definition. Even if the definition of the lifecycle is complete, there can still be uncertainty in terms of data quality; for instance the data about Part1 and Part2 in the material and distribution phases may be uncertain, resulting in data quality uncertainty. Even if the data quality is certain, there may still be uncertainty in methodological choices. For example, the method used for impact assessment may have been developed for a different region, is old or does not consider all the impact categories.

At any design stage, uncertainty in information available is a combination of these individual uncertainties. We need to identify what information is required in all these categories so as to accurately estimate the environmental impact of the product lifecycle at that stage and what information is available in all these categories at that stage; based on these, the uncertainty in impact estimation is assessed.

#### 4 Method Development

A method is developed using interval algebra and weighted objectives, which takes uncertainties about the product structure definition, lifecycle definition and data quality into account while assuming that the uncertainty related to methodological choices remains unchanged. This is because estimation of impact is always based on a particular methodology, and the uncertainty related to methodology will be the same for all proposals compared using that methodology.

During design, information about life cycle processes range from no selection (i.e., complete lack of data) to class selection (i.e., noncrisp data) to point selection (i.e., complete data). If we use probability theory [35], we need to have probability densities from previous data which is not available in LCI databases. Even if this data were available, this could be used only for crisp values, and not for noncrisp data such as classes as prevalent in the situations considered in this work. Dempster–Shafer theory [36] can

be used for sets (i.e., classes) but will require computation of belief and mass functions for each such class based on previous data, which is not available in LCI databases.

As a result, noncrisp data such as those corresponding to classes are represented in our method as intervals, which provide the range within which the value for the class should lie. Aggregation of such data from the life cycle processes, each with different impacts representing their relative importance, as required for LCA during earlier stages of design, require a method that integrates these data taking into account the relative importance. Development of a method that blends interval algebra and weighted objectives is a reasonable choice, therefore, for impact and uncertainty estimation in these situations. The proposed method offers an estimate of the environmental impact of a product lifecycle proposal as it evolves during various design stages while also providing an estimate of the uncertainty associated with the estimated impact in terms of a confidence (discussed below) on the impact estimated.

The proposed method has two major parts: impact estimation and uncertainty estimation. Impact estimation makes straightforward use of interval algebra—an established mathematical tool to deal with noncrisp values. Uncertainty estimation is harder. The challenge is to aggregate uncertainties associated with a list of processes, which fall into the following three categories of processes:

- having given impacts and uncertainty, both as intervals
- those that show no impacts as they have not been chosen by the designer but are known to exist
- those that have no impacts because they are not harmful to the environment

For aggregation, weighted objectives method is a commonly used Ref. [13] when criteria have different weights. In our case, the challenges of using weighted objectives are as follows.

- Impacts can be crisp or noncrisp values, and weights are proportional to the size of impact.
- Some processes cannot have weights since their impact values are zero by choice or by virtue of them being environmentally benign.

Our method uses a weighted sum on interval values by integrating weighted objectives method with interval algebra. Since both these are standard mathematical tools for decision making and are integrated in a manner ensuring that each applies to its designed domain of application, the method has a clear mathematical foundation. The processes that have zero values are counted in a non-weighted manner since weighting does not apply in these cases.

The method can be used to estimate, as an interval of values, the environmental impact of each chosen class or instance of a lifecycle process, for a given product as a collection of individual assemblies, subassemblies, parts, and interfaces. The method can then be used to aggregate these process-specific impacts into an overall impact measure for the product for its whole life cycle. Finally the method can be used to estimate the confidence on the impact of each individual process, and aggregate these to estimate the confidence on the overall impact of the product lifecycle.

The measure developed enables the impact value for a given class of lifecycle processes with given environmental impacts to be taken as an interval between two impact values—the maximum and the minimum possible in that class. The confidence level of an estimate is described using a number between 0 and 1, where 0 specifies no confidence on the estimation while 1 specifies 100% confidence. If for an entity (i.e., a part or an interface) neither a class nor a specific value is chosen for a given lifecycle phase (e.g., material phase), its impact is taken to be 0 with confidence equivalent to zero. If, on the other hand, any choice is made, confidence on the value of chosen is taken to be 1, which needs to

Table 2 Temporal, spatial relevance, and sample size

Years	Factor	Location	Factor	Sample size	Factor
<5	1	Country	1	Multiple	1
>5 & <10	0.94	Continent	0.98	Single	0.9
>10	0.88	Other continent	0.82		

408 be multiplied by the temporal factor, spatial factor, and sample  
 409 size factor (from Table 2) to account for the associated data un-  
 410 certainty.

411 Estimation of impact and confidence of a life cycle process is  
 412 performed as follows, for the four choices possible (the first two  
 413 referred henceforth as “zero-impact values,” while the remaining  
 414 two as “nonzero-impact values”).

415 1. No lifecycle processes are selected

416 Impact value<sub>i</sub> = 0 (1)

417 Confidence<sub>i</sub> = 0 (2)

418 2. A lifecycle process is selected with impact being zero

419 Impact value<sub>i</sub> = 0 (3)

420 Confidence<sub>i</sub> = 1 (4)

421 3. A lifecycle process class is chosen

422 Impact<sub>i</sub> = [V<sub>min</sub>V<sub>max</sub>]<sub>i</sub> \* ∏<sub>j=1</sub><sup>n</sup> LCPP<sub>j</sub> \* ∏<sub>k=1</sub><sup>m</sup> PSEP<sub>k</sub> (5)

423 Confidence<sub>i</sub> = [(tf \* sf \* ssf)<sub>min</sub>(tf \* sf \* ssf)<sub>max</sub>]<sub>i</sub> (6)

424 Here, *n* is the number of LCPP, *m* is the number of PSEP,  
 425 [V<sub>min</sub>V<sub>max</sub>] is the impact values in range for a specific unit of  
 426 lifecycle process range, *tf* is temporal factor, *sf* is spatial, *sf*  
 427 is sample size factor, LCPP (*life cycle process parameters*):  
 428 depend on the lifecycle process (for example for transportation,  
 429 distance in km), PSEP: These are *product structure*  
 430 *element parameters* and depend on the elements of the product  
 431 structure (e.g., part mass in kg). So for transporting a  
 432 product of *x* kg over *y* km, *x* and *y* need to be multiplied.  
 433 The final value is *x*\**y* kgkm, which is multiplied by the unit  
 434 impact value (specified in number of impact units per kgkm)  
 435 to estimate the impact of transportation of this product.

436 4. A specific lifecycle process is chosen

437 Impact value<sub>i</sub> = V<sub>i</sub> \* ∏<sub>j=1</sub><sup>n</sup> LCPP<sub>j</sub> \* ∏<sub>k=1</sub><sup>m</sup> PSEP<sub>k</sub> (7)

438 Confidence<sub>i</sub> = tf<sub>i</sub> \* sf<sub>i</sub> \* ssf<sub>i</sub> (8)

439 Here, V<sub>i</sub> is the impact value for a specific unit of lifecycle  
 440 process. Note that the specific values of these factors can  
 441 sometimes be derived from the analysis of life cycle inven-  
 442 tory data, such as those in Simapro databases [21]. The da-  
 443 tabase contains sets of data for each process; each data dif-  
 444 fers in terms of the time, space and the number of samples  
 445 from which it was created.

446 Depending on which data are picked for impact estimation and  
 447 which data best represent the time or space of the life cycle of a  
 448 product, an error will occur in the estimation that will vary from 0  
 449 to some absolute maximum value, depending on the choice of  
 450 data. The absolute mean percent error %e<sub>m</sub> for a given data set  
 451 representing a given process should be calculated as the average,  
 452 across all data-points in the set (extended from mean deviation in  
 453 statistics [37]), of the percent difference between the value of each

data point and the mean value of the data set; (1-%e<sub>m</sub>) is used as  
 the spatial or temporal factor depending on the nature of the data  
 set.

$$\%e_m = \frac{1}{n} \sum_{i=1}^n \left[ \frac{|v_k - v_m|}{v_m} \right]_i \quad (9)$$

Here, *n* is the number of data points, V<sub>k</sub> is the value of *k*th data  
 point, and V<sub>m</sub> is the mean value of the data set

For temporal relevance, the data sets available in the databases  
 consulted [17] are either within a 5 years span, or within a 10  
 years span. Figure 3 shows the absolute mean percent error (%e<sub>m</sub><sup>t</sup>)  
 for different processes, for 5 years and 10 years span. The average  
 %e<sub>m</sub><sup>t</sup> across all processes analyzed for a 5 year span is 6.13; for 10  
 year span it is 12.07; this implies that if older data is used, the  
 error increases. In these cases, the average temporal factors are  
 0.94 and 0.88, respectively.

For spatial relevance, typical data sets available in the databases  
 are either within a continent, or across continents. Figure 4 shows  
 the absolute mean percent error (%e<sub>m</sub><sup>s</sup>) in impact values for vari-  
 ous processes, plotted for data from the same continent and from  
 different continents. The average %e<sub>m</sub><sup>s</sup> across different processes  
 within a continent is 2; across continents it is 18 (nine times);  
 %e<sub>m</sub><sup>s</sup> is thus smaller within a continent than across continents. In  
 these cases, the average spatial factors are 0.98 and 0.82, respec-  
 tively.

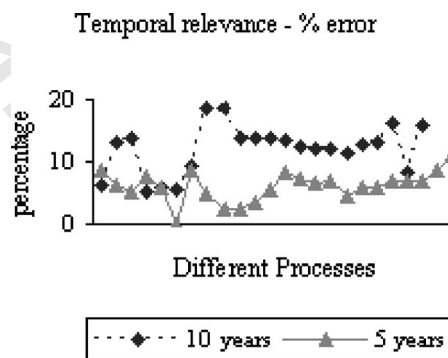


Fig. 3 Temporal relevance

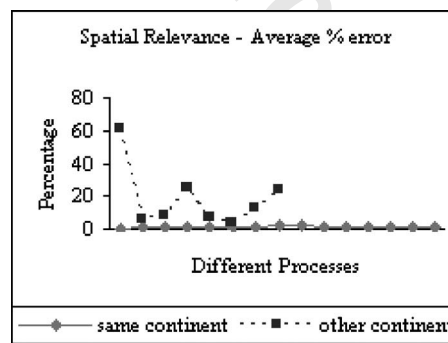


Fig. 4 Spatial relevance

477 Generally, data accuracy would be more if multiple samples are  
 478 used to create the data. For estimating the sample size factors, we  
 479 need the original samples from which each data-point in the da-  
 480 tabases, typically the average of the sample values, have been  
 481 created. The percent error  $\%e_s$  due to sample size variation is  
 482 calculated using Eq. (10) [37]. However, current databases do not  
 483 provide these individual samples. As placeholders, we currently  
 484 take 1 for multiple data and 0.9 for single data; in reality the  
 485 database provider can be asked to provide the original samples so  
 486 that the sample size error can be accurately estimated using Eq.  
 487 (10).

488 
$$\%e_s = \left[ \frac{t * s}{\sqrt{n}} \right] \quad (10)$$

489 Here,  $n$  is the number of samples, which is 1,  $s$  is the standard  
 490 deviation, and  $t$  is the factor based on  $n$  should be taken from  
 491 t-table.

492 Based on the above discussion, indicative average temporal fac-  
 493 tor, spatial factor and sample size factors are provided in Table 2.  
 494 For instance, for data less than 5 years old, the temporal factor  
 495 could be taken as 1, i.e., 100% accurate; for data older than 5  
 496 years and less than 10 years, the factor could be taken as 0.94, and  
 497 so on. For the Spatial values, if the data is from the same country,  
 498 the factor could be taken as 1; if it is not from the same country  
 499 but from the same continent, the factor could be taken as 0.98, and  
 500 so on. According to the sample size values in the Table 2, if the  
 501 data are from multiple samples, the factor is taken as 1, if it is  
 502 from a single source, the factor is taken as 0.9. With greater data  
 503 availability, these values could be made more specific to the pro-  
 504 cess, space, time and sample size used. In the example case dis-  
 505 cussed in scenario 1 (Sec. 6), let the minimum value of a process  
 506 class in the material phase be temporally within 5 years and spa-  
 507 tially across continent from the usage scenario, and based on mul-  
 508 tiple samples; the maximum value for this process class be tem-  
 509 porally over 5 years, spatially within continent, and is also based  
 510 on multiple samples; then the confidence interval of the process  
 511 class, estimated using Eq. (6), is  $(1 \times 0.82 \times 1.94 \times 0.98 \times 1)$   
 512  $\sim (0.8 \quad 0.9)$ .

513 **4.1 Lifecyclewise Impact and Confidence Estimation.** The  
 514 impact of a product in the material phase is an aggregation of the  
 515 individual material impacts of its assemblies, subassemblies and  
 516 parts. Equation (11) is used to estimate the impact of a product in  
 517 the material phase.

- 518 • The impact of an assembly in the material phase is an ag-

- gregation of the individual material impacts of the assem- 519  
 blies, subassemblies and parts in that assembly. 520  
 • The impact of a subassembly in the material phase is an 521  
 aggregation of the material impacts of the parts in that sub- 522  
 assembly; interfaces have no material impact. 523  
 • The impact of a part in the material phase is an aggregation 524  
 of the impacts of the individual material processes in that 525  
 part. 526

Equation (12) is used to estimate the confidence on the impact 527  
 of a product at the material phase. The aggregated confidence on 528  
 the impact value is dependent on the impact values (as weights) 529  
 and the associated confidences for nonzero impact values, and 530  
 only on the confidences for zero impact values (since zero-impact 531  
 values have no impact). This is to take into account the fact that 532  
 our confidence of a sum of values is affected proportionately by 533  
 the values as well as confidence on them: a given confidence on a 534  
 larger value influences more the confidence on the sum than does 535  
 the same confidence on a smaller value. 536

Similar logic is used for finding the impacts and associated 537  
 confidences for other subsystems and other lifecycle phases. 538

The assumptions behind Eqs. (11) and (12) are the following. 539

- Impact in a given life cycle phase can be estimated by ag- 540  
 gregating the impacts of all processes in that phase. 541
- For a life cycle process, the impact will be zero (if it is 542  
 environmentally benign) or nonzero (if not). In each case, 543  
 there will be some confidence on this impact. 544
- The confidence on the aggregate value (for a given phase) 545  
 will be shared proportionately by the aggregate confidence 546  
 on the zero-impact value processes and the nonzero-impact 547  
 value processes. 548
- The aggregate confidence of the nonzero-impact value pro- 549  
 cesses is proportional to the number of these processes as 550  
 well as the value and confidence of the processes. 551
- The aggregate confidence of the zero-impact value pro- 552  
 cesses is proportional to the number of these processes as 553  
 well as the confidence on each such process (since the im- 554  
 pact value is zero in these cases). 555
- For normalization purposes, the equation should reflect that 556  
 the scale of aggregate confidence on the impact value within 557  
 a phase should be between 0 and 1. 558

559 
$$PI_iM = \sum_{j=1}^{\text{No. of } A} AI_{ij}M + \sum_{l=1}^{\text{No. of } SA} SAI_{il}M + \sum_{n=1}^{\text{No. of } Pa} PaI_{in}M \quad (11)$$

560  
561

562

563

564

565

566

567

$$PC_iM = \frac{NZ_i}{NZ_i + Z_i} \left[ \frac{\sum_{j=1}^{NZ_{iA}} AI_{ij}M * AC_{ij}M + \sum_{l=1}^{NZ_{iSA}} SAI_{il}M * SAC_{il}M + \sum_{n=1}^{NZ_{iPa}} PaI_{in}M * PaC_{in}M}{\sum_{j=1}^{NZ_{iA}} AI_{ij}M + \sum_{l=1}^{NZ_{iSA}} SAI_{il}M + \sum_{n=1}^{NZ_{iPa}} PaI_{in}M} \right] + \frac{Z_i}{NZ_i + Z_i} \left[ \frac{\sum_{a=1}^{Z_{iA}} AC_{ia}M + \sum_{b=1}^{Z_{iSA}} SAC_{ib}M + \sum_{c=1}^{Z_{iPa}} PaC_{ic}M}{\sum_{a=1}^{Z_{iA}} AC_{ia_{max}}M + \sum_{b=1}^{Z_{iSA}} SAC_{ib_{max}}M + \sum_{c=1}^{Z_{iPa}} PaC_{ic_{max}}M} \right] \quad (12)$$

$NZ_i = NZ_{iA} + NZ_{iSA} + NZ_{iPa}Z_i = Z_{iA} + Z_{iSA} + Z_{iPa}$



**Table 3 Impact and confidence of life cycle processes of product elements in different scenarios**

		Material phase		Production phase		Distribution phase		Usage phase		After-usage phase	
		$I_m$	$C_m$	$I_p$	$C_p$	$I_d$	$C_d$	$I_u$	$C_u$	$I_a$	$C_a$
S1	Part1	[2,4]	[0.8 0.9]	[2,3]	[0.9 1]	[1,2]	[0.3 0.4]	0	1	–	0
	Int1			–	0						
S2	Part2	[1,2]	[0.5 0.6]	2	1	[1,3]	[0.4 0.4]	0	1	–	0
	Part1	[2,4]	1	[2,3]	1	[1,2]	1	0	1	–	0
S3	Int1			–	0						
	Part2	[1,2]	1	2	1	[1,3]	1	0	1	–	0
S4	Part1	[2,4]	1	[2,3]	1	[1,2]	1	0	1	[2,3]	1
	Int1			–	0						
S4	Part2	[1,2]	1	2	1	[1,3]	1	0	1	1	1
	Part1	4	1	3	1	2	1	0	1	3	1
	Int1			2	1						
	Part2	2	1	2	1	3	1	0	1	1	1

568 Here,  $PI M$  is the product environmental impact in material  
 569 phase,  $AI M$  is the assembly environmental impact in material  
 570 phase,  $SAI M$  is the subassembly environmental impact  
 571 in material phase,  $PaI M$  is the part environmental impact in  
 572 material phase,  $M$  is the material,  $A$  is the assemblies,  $SA$   
 573 is the subassemblies,  $i$  is the identifier for product,  $j$  is the  
 574 identifier for assembly,  $l$  is the identifier for subassembly,  $n$   
 575 is the identifier for part,  $PC M$  is the product confidence in  
 576 material phase,  $AC M$  is the assembly confidence in material  
 577 phase,  $SAC M$  is the subassembly confidence in material  
 578 phase,  $PaC M$  is the part confidence in material phase,  $NZ_i$   
 579 is the total number of nonzero valued items in material phase in  
 580 product  $i$ ,  $NZ_A$  is the number of nonzero valued  
 581 assemblies,  $NZ_{SA}$  is the number of nonzero valued sub-  
 582 assemblies,  $NZ_{Pa}$  is the number of nonzero valued parts,  $Z_i$  is  
 583 the total number of zero valued items in material phase in  
 584 product  $i$ ,  $AC_{max} M$  is the maximum possible Assembly con-  
 585 fidence in material phase,  $SAC_{max} M$  is the maximum pos-  
 586 sible subassembly confidence in material phase,  $PaC_{max} M$   
 587 is the maximum possible part confidence in material phase,  
 588  $Z_A$  is the number of zero valued assemblies,  $Z_{SA}$  is the num-  
 589 ber of zero valued subassemblies, and  $Z_{Pa}$  is the number of  
 590 zero valued parts.

591 **4.2 Overall Impact and Overall Confidence.** The method  
 592 for estimating the overall impact value and the overall confidence  
 593 on this impact for a product lifecycle is based on Eqs. (13) and  
 594 (14), with similar assumptions as in Eqs. (11) and (12) but now  
 595 applied to zero-impact-value and nonzero-impact-value life cycle  
 596 phases (rather than processes).  
 597 The estimates on impact and associated confidence of product  
 598 structure elements, life cycle phases, etc. will be aggregated to-  
 599 gether to form the impact of the overall product. There are two  
 600 possible levels of addition: addition of impacts of all the child  
 601 elements (e.g., extraction, production, and distribution) in a parent  
 602 element for a given life cycle phase (e.g., material), and addition  
 603 of impacts from all life cycle phases. The addition of impacts is  
 604 carried out using interval algebra while estimation of confidence  
 605 is made using a weighted sum of the individual confidence of  
 606 impacts, where the impact values are used as the weights.  
 607 The overall environmental impact of the product lifecycle pro-  
 608 posal is estimated by adding all nonzero individual lifecycle im-  
 609 pacts, using Eq. (13). The overall confidence is estimated by tak-  
 610 ing both nonzero-impact lifecycle phases and zero-impact  
 611 lifecycle phases, using Eq. (14).

$$612 \quad PI_i \text{ total} = \sum_{l=1}^{\text{No. of LCP in } i\text{th product}} PI_{il}LCP \quad (13)$$

$$PC_i \text{ total} = \frac{NZ_L}{NZ_L + Z_L} \left[ \frac{\sum_{j=1}^{NZ_L} V_{ij} * C_{ij}}{\sum_{j=1}^{NZ_L} V_{ij}} \right] + \frac{Z_L}{NZ_L + Z_L} \left[ \frac{\sum_{k=1}^{Z_L} C_{ik}}{\sum_{k=1}^{Z_L} C_{ik_{max}}} \right] \quad (14) \quad 613$$

614 Here  $PI$  total is the overall product environmental impact (in all  
 615 lifecycle phases),  $PI$  LCP is the product environmental impact  
 616 (lifecycle wise),  $PC$  total is the overall product confidence (in all  
 617 lifecycle phases),  $NZ_L$  is the number of nonzero valued lifecycle  
 618 phases in  $i$ th product,  $Z_L$  is the number of zero valued lifecycle  
 619 phases in  $i$ th product,  $V_{ij}$  is the environmental impact values of  $j$ th  
 620 lifecycle phase of  $i$ th product,  $C_{ij}$  is the confidence of  $j$ th lifecycle  
 621 phase of  $i$ th product,  $C_{ik}$  is the confidence of  $k$ th lifecycle phase of  
 622  $i$ th product, and  $C_{ij_{max}}$  is the maximum confidence of  $k$ th life-  
 623 cycle phase of  $i$ th product. For a range of values of  $V_i$ , we get  
 624 confidence in range

**5 Calculation Example** 625

626 The proposed method for assessing environmental impact and  
 627 associated confidence is evaluated using a set of example sce-  
 628 narios with varying levels of uncertainty. The hypothesis is that,  
 629 those scenarios that have more elements in the various categories  
 630 of uncertainty that are not considered, known, known precisely, or  
 631 known with what relevance, are likely to have less environmental  
 632 impact estimates with less confidence values. As the number of  
 633 such elements is reduced, the impact values should increase with  
 634 an associated increase in their confidence.

635 **AQ: #3**  
 636 Let us take the example from Sec. 4.1 in which a product pro-  
 637 posal has two individual parts Part1 and Part2 with one interface  
 638 Int1; the impact values (in intervals) and confidence on these val-  
 639 ues (also in intervals) for the parts and the interface, in various life  
 640 cycle phases, are specified in Table 3. Here,  $I$  is used to denote  
 641 impact value and  $C$  to denote confidence. Five lifecycle phases are  
 642 considered: material, production, distribution, usage and after-  
 643 usage.

644 Four scenarios are taken to check the consistency of calculation  
 using the above equations.

- 645 • Scenario1 (S1): Uncertainty exists in all three categories  
 646 (product-structure, lifecycle, and data quality). Data on Int1  
 647 is not available (reflected in no values in impact or uncer-  
 648 tainty for Int1 in the table), which accounts for uncertainty  
 649 in product structure, the after-usage details of the parts are  
 650 not specified, which gives uncertainty in lifecycle, and the

Table 4 LCP wise overall impact values and associated confidence in the four scenarios

	Material phase		Production phase		Distribution phase		Usage phase		After-usage phase		Over all	
	I <sub>m</sub>	C <sub>m</sub>	I <sub>p</sub>	C <sub>p</sub>	I <sub>d</sub>	C <sub>d</sub>	I <sub>u</sub>	C <sub>u</sub>	I <sub>a</sub>	C <sub>a</sub>	I	C
S1	[3,6]	[0.3 1]	[4,5]	[0.5 0.6]	[2,5]	[0.2 1]	0	1	0	0	[9,16]	[0.3 0.8]
S2	[3,6]	[1]	[4,5]	[0.5 0.6]	[2,5]	[1]	0	1	0	0	[9,16]	[0.4 0.8]
S3	[3,6]	[1]	[4,5]	[0.5 0.6]	[2,5]	[1]	0	1	[3,4]	[1]	[12,20]	[0.6 1]
S4	6	1	7	1	5	1	0	1	4	1	22	1

651 initial confidence values of the parts in material and distri-  
 652 bution phases are uncertain, resulting in data quality uncer-  
 653 tainty.  
 654 • Scenario2 (S2): Uncertainty exists in two categories (prod-  
 655 uct structure and lifecycle). Here also, the product structure  
 656 and lifecycle uncertainty are the same as in S1 but data  
 657 quality uncertainty is removed by selecting relevant data, as  
 658 seen in the new confidence value of 1.  
 659 • Scenario3 (S3): Uncertainty exists in only one category  
 660 (product structure). Here, the product structure uncertainty  
 661 remains as before while the lifecycle uncertainty is removed  
 662 by specifying the necessary values for Part1 and Part2 in the  
 663 after-usage phase.  
 664 • Scenario4 (S4): There is no uncertainty: all the required data  
 665 is available. Here, necessary values for Int1 are provided, so  
 666 that the product structure uncertainty is removed and thus  
 667 the overall confidence should be 1.

668 The four scenarios in the example are designed such that the  
 669 amount of information about the product lifecycle proposal is in-  
 670 creased steadily from scenario 1 to scenario 4. If the proposed  
 671 method for assessing impact and associated uncertainty is reason-  
 672 able, it should predict a steady increase in the impact value and a  
 673 steady reduction in uncertainty reflected in a steady increase in  
 674 confidence. For it to be acceptable, the method embodied in the  
 675 equations should be able to provide estimates on the confidence in  
 676 the calculated impact value as would be intuitively expected in the  
 677 scenarios, which are varied depending on the lack of detailed  
 678 information about the product lifecycle proposal. Table 4 shows  
 679 the summary of impact values and the estimates of confidence on  
 680 them, as estimated using the proposed method. As can be seen  
 681 from this table, as the uncertainty is reduced across the scenarios,  
 682 the impact values and confidences on them are increased as ex-  
 683 pected. In S1, the overall impact ranges from 9 to 16; the confi-  
 684 dence on this impact ranges from 0.33 to 0.8. The difference be-  
 685 tween S1 and S2 is only in data uncertainty; so the overall impact  
 686 value remains the same (9 16) but the lower value of the confi-  
 687 dence range increases (0.46 0.8).

688 From S2 to S3, further information about the after-usage phase  
 689 is added, resulting in an increase in both impact value (12 20) and  
 690 confidence (0.6 1). In S4, information about product structure el-  
 691 ement is also added, resulting in an increase in both impact value  
 692 (22) and confidence (1). Note that during design, multiple life  
 693 cycle alternatives may have to be compared, taking into account  
 694 both impact and associated confidence. For example, if an alter-  
 695 native has a greater impact with greater confidence than another,  
 696 choice using traditional methods will favor the latter for its lower  
 697 impact. This might be an error in judgment since the impact esti-  
 698 mated for the latter is less complete, as shown in its lower confi-  
 699 dence, and hence likely to increase more as more information  
 700 becomes available; the decision might have to be deferred. For  
 701 these situations, new decision methods are necessary. One such  
 702 method is discussed in Ref. [38].

703 **6 Summary and Conclusions**

704 Various categories of uncertainty associated with using LCA for  
 705 a product lifecycle proposal in various stages of design are iden-  
 706 tified. Based on this, a method for estimating lifecycle environ-

mental impact and associated confidence of a product is devel-  
 707 oped. This method is evaluated using example scenarios with  
 708 varying uncertainty. 709

As the method is capable of taking into account all three cat-  
 710 egories of uncertainty, it is likely to be better suited to support  
 711 decision-making throughout the design process where information  
 712 continues to develop and uncertainties progressively get reduced. 713  
 The scope for this paper is using LCA in design for estimating  
 714 impacts and associated uncertainties. Within this, methodological  
 715 uncertainty is currently not addressed. Also, for a different meth-  
 716 odology such as MET matrix, the nature of uncertainties might be  
 717 different. These need further investigation. 718

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- #1 Au: References 26 and 31 are the same. Please check our renumbering of Refs. 31–38.
- #2 Au: Please provide the definition of CAD,

CATIA, LCP, and MET if possible.

- #3 Au: Please check change from Sec. 4.5 to Sec. 4.1 as there is no Sec. 4.5 in this paper.

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